

University of South Bohemia in České Budějovice

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**Edge of habitability: bird community on dump sites after
uraninite mining**

Master Thesis

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Annotation

Biota on post-mining areas is a well-examined topic in restoration ecology. However, most scientific attention focuses on large-scale opencast mining sites or spoil heaps. This study attempts to offer an insight into ecological conditions on a unique ecosystem represented by dump sites after uranium mining, using birds as a study group. We examined differences in bird abundance, species richness, community structure and habitat preferences between nine moderately sized dump sites and eight control areas. Our study can function as a pilot survey of birds of uranium mining sites, offering a broad overview of population-level effects of hostile conditions. Our findings can guide restoration practices with respect to this specific type of ecosystem.

I hereby declare that the submitted Master's degree thesis was written solely by me without any third-part assistance, information other than provided sources or aids have not been used and those used have been fully documented.

Keywords

Birds, post-industrial site, uranium mining, community structure, vegetation characteristics

In České Budějovice 7.11.2023

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Bc. Kryštof Korejs

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Declarations

This thesis is part of a publication in preparation entitled: Edge of habitability: bird community on dump sites after uraninite mining. The author of this thesis is also the primary author of the publication, and the corresponding author. The corresponding author confirms that this publication was also co-authored by doc. Mgr. Jan Riegert, Ph. D., and Mgr. Vojtěch Dolejšek. Study and survey design was performed together by primary author (50%) and Jan Riegert (50%). Majority of data collection was performed by the primary author (90%) and was supplemented by Vojtěch Dolejšek (10%), who assisted with the vegetation survey. Statistical analyses were performed, and first draft of manuscript was written by the primary author (100%). Subsequent edits were performed together by the primary author (50%) and by Jan Riegert (50%). All authors read and approved the final manuscript.

Manuscript

Title: **Edge of habitability: bird community on dump sites after uraninite mining**

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Abstract

Biota on post-mining areas is a well-examined topic in restoration ecology. However, most scientific attention focuses on large-scale opencast mining sites or spoil heaps. This study attempts to offer an insight into ecological conditions on a unique ecosystem represented by dump sites after uranium mining. Birds were used as a study group during field monitoring. We gathered data on differences in abundance, species richness, community structure and bird habitat preferences between nine moderately sized dump sites and eight control areas. Vegetation structure and invertivore food supply were also examined. Statistical analyses included constrained ordinations and mixed-effect GLMs. Dump sites attracted valuable bird assemblages despite extreme soil conditions, increased levels of radioactivity near ground, limited development of vegetation, and an impoverished food supply. These assemblages included primary succession specialists, declining grassland species, nomadic granivores or perching raptors. However, overall bird abundances as well as species richness were lower compared to control sites, and some species had to significantly adjust their ecological requirements for their survival. Radioactivity near exposed pieces of uraninite rocks could impact living conditions of ground-dwelling bird species, together with absorption of radionuclides by insects or vegetation. In addition, comparisons with other studies on birds in restoration science reveal key differences from opencast mining sites. Our study can function as a pilot survey of birds of uranium mining sites, offering a broad overview of population-level effects of hostile conditions. Our findings can guide restoration practices with respect to this specific type of ecosystem.

List of Abbreviations

RDA – Redundancy analysis

CCA – Canonical correspondence analysis

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1 **1. Introduction**

2

3 As habitat loss draws many organisms to the fringes of existence, post-industrial areas are
4 becoming increasingly important for biodiversity conservation. Abandoned mines, quarries and
5 spoil heaps are often the largest sites impacted by past industrial activity (Navarro-Ramos et al.
6 2022). Most of the world's mining land use exists within vast, resource-rich countries such as
7 the Russian Federation, China, or Australia, but it is also prevalent in European countries such as
8 Germany, Czech Republic or Ukraine (Maus et al. 2022). These countries have a long history of
9 industrial development, with much of their landscapes still dealing with the ecological impacts
10 of past mineral extraction (Krümmelbein et al. 2012). In Central Europe, these impacts are
11 generally addressed either by large-scale modification by heavy machinery followed by assisted
12 revegetation (Hendrychová 2008), or by creation of artificial water reservoirs in place of mines
13 (Molenda & Kidawa 2020) or by leaving areas to spontaneous succession (Tropek et al. 2012).

14 While the last approach may be suitable for maximizing biodiversity services (Tropek et
15 al. 2010), it is often applied to areas where extreme conditions impede plant and animal
16 colonization (Prach & Hobbs 2008). In such cases, ecosystem succession tends to proceed slowly
17 and to a limited extent. First plant settlement is represented by pioneer grasses and forbs with
18 very low vegetation cover (Skousen et al. 1990). Tree and shrub species follow after several years,
19 slowly facilitating a perfunctory canopy layer (Skousen et al. 1990). In Central Europe, they are
20 represented by hardy colonizers reliant on anemochory (*Betula* sp., *Salix* sp., Mudrak et al. 2010).
21 Presence of vegetation is most often impacted by thickness of the fermentation layer. However,
22 expansive grasses can arrest succession in nutrient-rich areas by outcompeting other understory
23 species (e.g., *Calamagrostis epigejos*, Mudrak et al. 2010). Thickness of fermentation layer is
24 closely related to the overall soil quality. Post-mining sites that do not undergo assisted
25 revegetation or surface modification are often located on poor substrates with high concentration
26 of basic elements (Vesela et al. 2021), poor water retention capacity (Cejpek et al. 2013) and an
27 increased proportion of pollutants and toxic compounds (Frouz & Vinduřkova 2018). In addition,
28 many unmodified sites offer little protection from weather phenomena, with high temperature
29 variation between night and day, exposure to wind, rainfall and consequent soil erosion (Prach &
30 Hobbs 2008). Similar to plants, animal colonization is impacted by these hostile conditions, and
31 primary succession specialists tend to dominate (Bejcek & řastny 1984, řalek 2012).

32 A less-studied example of post-mining areas that develop extreme conditions are dump
33 sites created by underground mining. These are mostly by-products of extraction of hard coal,
34 ores, but also uraninite rock (Sasková 2011). Underground mining has a much lower overall
35 ecological footprint than open-pit mines, since it does not directly strip the landscape. On the
36 other hand, it produces coarse waste products (slag and tailing) that are deposited on the surface
37 (Blight 2011). The resulting dump sites consist of steep heaps of rocks and mining slag,
38 combining all previously listed extreme conditions, often with high content of toxic compounds
39 (Peterková 2021). However, in case of some types of underground mines, there is an additional
40 unique effect of radioactivity. Specifically, radioactivity associated with mining of thorium, or
41 more often, uranium ores (Jull et al. 1987). This is to distinguish uranium mines from mining
42 sites that are affected by unrelated “natural” radioactivity levels (Darko et al. 2010). Radioactivity
43 associated with extracting these ores has been long under scrutiny, and many protective measures
44 exist to reduce environmental contamination (Robinson 2004). Despite this, radiation levels are
45 often increased at most sites of extraction, especially in reservoirs of runoff water and
46 neutralization ponds (Carvalho et al. 2011). The coarse dump site material often contains residual
47 uranium in pieces of leftover ore, facilitating increased radioactivity levels (Carvalho et al. 2007).
48 Resulting exposure to radiation can impact organisms in abandoned mining areas (Hinck et al.
49 2017). Despite this, biodiversity on former uranium mining sites has not been sufficiently studied.
50 In fact, effects of radioactivity on wildlife are much more often examined on landscapes irradiated
51 by nuclear accidents (Wehrden et al. 2012). The associated high levels of radiation are affecting
52 large areas and impact the environment more severely than radiation associated with mining
53 (Friedman et al. 1987). However, findings of studies within these irradiated areas can be
54 applicable to post-mining sites, albeit at a smaller scale.

55 To narrow down the available information, we will focus on bird response to increased
56 radioactivity. Studies on birds from the Chernobyl area have shown a significant decrease in
57 sperm count (Moller et al. 2014), brain size (Møller et al. 2011), and plumage quality (Møller et
58 al. 2013), but also decreases in community metrics such as bird abundance or diversity, which
59 were observed at a similar scale in Fukushima (Møller et al. 2012). Another source of
60 radioactivity can be associated with nuclear bomb tests. A review by Mellinger et al. (1975) has
61 discussed increased presence of radionuclides in the internal digestive tracts of birds as early as
62 three years after the Trinity test (1948), and contaminated white-fronted geese (*Anser albifrons*)
63 populations in Northern Europe after Soviet nuclear tests (1962). Indeed, in comparison with
64 these far-reaching sources of radiation exposure, mining-related radiation is a much weaker

65 effect, but nevertheless capable of altering living conditions for animal life (Hinck et al. 2017)
66 and therefore worthy of study.

67 To conclude, there exist sample literature resources concerning the effects of radioactivity
68 reviewed above, but these effects are mostly related to large-scale nuclear events. Concurrently,
69 while surface post-mining areas are well-examined by ecologists, no attention is paid to dump
70 sites after extraction of uranium. This twofold disregard can be addressed by a study that
71 investigates ecological conditions on these sites. This can be achieved by choosing a group of
72 organisms and gauging their response to on-site conditions. Birds can be most suitable for
73 multiple reasons. They respond quickly to environmental changes and are relatively easy to
74 monitor (Helms et al. 2018). At the same time, as primary, secondary, or even tertiary consumers
75 within trophic networks, they are likely to be affected by accumulation of toxic or radioactive
76 contaminants that may impact community metrics (Weeks et al. 2022). Lastly, enough literature
77 exists concerning birds on non-radioactive post-mining sites (Hendrychová 2008), which can
78 allow for comparisons.

79 As most studies on birds on post-mining sites were performed on Central-European spoil
80 heaps (e.g., Šálek 2012, Moudrý et al. 2021, Korejs et al. 2023), choosing a study site from this
81 region may increase comparability of results. Therefore, we selected an area with a rich history
82 of underground uranium extraction, the surroundings of the city of Příbram in the southern part
83 of Central Bohemia. The first mines became active in 1952 (Sasková 2011). Due to mining
84 activities, large amounts of slag and tailing have been accumulated over a 40-year period of
85 intensive excavations, forming large and steep dump sites. By the end of the Cold War, mining
86 operations began to slow down, and vegetation succession began to proceed on abandoned
87 dumping sites (Peterková 2021). Then the sites were transferred under the jurisdiction of the
88 Czech governmental organization DIAMO, responsible for management of irradiated post-
89 industrial areas (Šimáčková 2017, DIAMO State Enterprise 2023). These dump sites are
90 comparable to other nutrient-poor, rocky, originally barren mines and spoil heaps with no surface
91 modifications (Prach & Hobbs 2008). However, there is an additional component of radiation.
92 While the ambient radiation levels are negligible, small amounts of uranium rock exist on the
93 surface that can impact animal and plant life (Samuel-Nakamura 2013, author observations, see
94 Methods).

95 This study was performed to expand known information on a set of under-investigated
96 topics. As was reviewed above, post-mining areas after uranium extraction are not sufficiently
97 examined by ecologists, and effects of radioactivity in general are mostly studied in relation to

98 large events such as weapons testing or nuclear accidents. The dump sites in surroundings of
99 Příbram are suitable study sites for examining both the general effects of dump sites on bird
100 communities and to make comparisons other post-mining sites. The main aims were to 1)
101 examine differences in biotic conditions such as vegetation parameters or invertebrate food
102 supply during a preliminary survey, 2) explore how do bird community metrics (abundance,
103 species richness, community structure) differ between dump sites and the surrounding landscape,
104 3) compare habitat preferences between birds on dump sites and surrounding landscape, and
105 finally 4) discuss differences in community metrics between dump sites after uranium mining
106 and post-mining sites in general, as well as the conservation potential of dump sites as habitats
107 for bird species. Based on these findings, this study can function as a pilot survey for further
108 research on birds of radiation-contaminated mining areas. In addition, results on bird community
109 response to extreme conditions can reinforce existing knowledge on ecological management of
110 post-mining sites.

111

112 **2. Methods**

113

114 The study took place on 17 survey sites located near the city of Příbram in Central Bohemia
115 (Figure 1). We surveyed nine dumping sites after uranium mining, while eight areas in the
116 surrounding landscape functioned as control sites. The dump sites were of varying sizes (total
117 area: 52 ha; mean area \pm s.d.: 6.5 ± 5 ha), and all contained mining tailing and slag with residual
118 amounts of uranium dumped in different periods during the last century (Supplementary material,
119 Table S1). Control sites were of similar sizes as dump sites (total area: 68 ha; mean area \pm s.d.: 8
120 ± 6 ha). They featured planted forests dominant in surrounding cultural landscape (Lipský 2000),
121 interspersed with low-productivity grass patches. While the overall radioactivity levels measured
122 at these sites are reported to not be life-threatening (Morávková 2006), these can increase near
123 exposed pieces of uraninite rock, containing residual uranium. Using a Geiger-Müller radiometer,
124 we found that radioactivity levels emitted by pieces of this mineral reach at least 100 μ Sv per
125 hour, which is a similar radiation exposure as received during a full-body x-ray scan (Perko et al.
126 2015). Therefore, radioactivity may affect plant and animal life particularly on the ground near
127 exposed uraninite. In contrast, control sites were not impacted by radioactivity and were fully
128 integrated within the surrounding agricultural-forest mosaic. However, they can be still affected

129 by the industrial history of the landscape, as the Příbram area has been heavily reshaped to
130 accommodate its uranium mining infrastructure (Sasková 2011).

131 Because of the extreme conditions of the dump sites, vegetation cover was likely to be
132 different from control sites. These differences were quantified by a survey, which took place in
133 early August of 2023. During this survey, the cover of different vegetation layers was noted for
134 each study site and a 50-meter buffer zone around each mining site. Methodology on classifying
135 vegetation structure was taken from Šálek (2012). Herb layer corresponded to the proportional
136 cover of herbaceous plants growing directly above ground. Understory layer corresponded to the
137 proportional cover of woody vegetation as well as young trees up to approximately four meters
138 in height. Canopy layer cover was represented by the connectivity of the forest canopy taller than
139 approximately four meters. Information collected on vegetation layer cover was integrated with
140 ArcGIS Pro (Esri 2023). Existing bodies of vegetation were recorded as polygon features, while
141 noting down the proportional cover of each layer within their respective attribute tables as a
142 fraction. These fractions were multiplied by the total area of each respective polygon and divided
143 by the total area of each dump site and surrounding buffer zone. This way, final proportions of
144 vegetation cover for each layer were reached. We also classified the dominant woody plant
145 species for each study site by noting down their presence (0/1).

146 Another survey was conducted in early August 2023 to examine the differences in
147 invertebrate community on dumps and control sites. The purpose of this survey was to determine
148 whether there were significant differences in food supply for invertivorous bird species between
149 dump and control sites. The survey was conducted via beating of *Petula pendula* branches
150 reachable by hand for 30 minutes per each study site. Gathered invertebrates were then classified
151 into orders, and their abundances on each site were recorded.

152 The main survey examining bird community was performed in 2022 on dump sites, and
153 in 2023 on control sites. Birds were sampled for each study site exactly three times (late April,
154 late May and in late June). To include birds not directly present on the sites, but close enough so
155 their territories could encroach on them, we also sampled a 50-meter buffer zone around each
156 dump site. Each site was surveyed while walking slowly and covering the entirety of the site
157 interior as well as the buffering zone. All bird individuals exhibiting territorial behaviour (such
158 as song, alarm call or continued perching) were recorded. Maximal abundances of each species
159 on each study site and its surrounding buffer zones over all three visits were used for further
160 analyses. We also estimated the approximate centre of each bird's territory by observing their
161 movements during our visit and marked down its geographical coordinates. For birds whose

162 assumed territory centres were located directly on study sites and not within the buffer zones, we
163 created 50-meter buffers around each centre in ArcGIS Pro. Then by intersecting the 50-meter
164 buffers with existing polygon features delimiting vegetation cover within dump sites, we created
165 new polygon features with information on vegetation layer cover within each territory. The final
166 proportional covers were reached using the same methodology as in the main vegetation survey.
167 This was done to compare bird habitat preferences on dump and control sites by using each bird
168 territory on each visit as an independent observation in community structure analyses.

169 A one-tailed Mann-Whitney U-test (McKnight & Najab 2010) in R 4.3.5 software (R Core
170 Team 2023) was used to compare the cover of all vegetation layers on dump and control sites. In
171 addition, the dominant woody plant species on dump and control sites were used to perform a
172 redundancy analysis (RDA) in Canoco 5 software (Šmilauer & Lepš 2014). The purpose of this
173 analysis was to determine whether the plant community on dump sites differed from control sites,
174 and to show which species were more likely to be present on each type of the site. The response
175 variable was the presence/absence (1/0) of each dominant woody plant species recorded on sites,
176 while the explanatory variable (primary predictor) specified site type (dump/control site). The
177 RDA is a linear method of constrained ordination, which allows us to explore the proportion of
178 variability in our dataset (presence of dominant species) that can be explained by primary
179 predictors (site type), and simultaneously testing the null hypothesis that the perceived variability
180 in response data results from random processes rather than environmental effects (Zuur et al.
181 2007).

182 A second RDA analysis was performed to examine how the invertebrate food supply for
183 invertivorous birds differed between site types. We used the total number of individuals of each
184 invertebrate order gathered on sites as a response variable, and site type (dump/control site) as
185 explanatory variable.

186 The data gathered during the bird survey were used in multiple analyses, in which birds
187 from inside the 50-meter buffer zones entered together with birds from the respective study sites.
188 First, a one-tailed Mann-Whitney U-test (McKnight & Najab 2010) in R 4.3.5 software (R Core
189 Team 2023) was used to compare the number of bird species and total abundances of birds on
190 dump and control sites, with site type being the explanatory variable (dump/control site). Then a
191 canonical correspondence analysis (CCA) in Canoco 5 software was used to compare bird
192 community structure between dump and control sites. Abundances of different bird species were
193 used as dependent variables, while the site type was used as explanatory variable. The CCA

194 method follows the same rationale as the RDA but is more suitable for response data with non-
195 linear distribution (Šmilauer & Lepš 2014).

196 Lastly, habitat preferences of seven common bird species were examined by a set of
197 mixed-effect linear models using the glmmTMB package (Magnusson et al. 2017) in R 4.3.5
198 software. The response variables were represented by proportional covers of different vegetation
199 layers within the 50-meter radius around assumed bird territory centres. For each response
200 variable, a separate model was built: 1) herb layer cover, 2) understory cover, 3) canopy cover.
201 Two models (1, 3) were fitted using a negative binomial distribution with log link function, and
202 the remaining model (2) was fitted using a gaussian distribution and identity link function (Hardin
203 & Hilbe 2007). Each model used a single independent variable (fixed effect): site type
204 (dump/control site), and two random effect variables: bird species to differentiate between
205 different birds' territories and consequent preferences of vegetation, and study site id to account
206 for the similarity in vegetation cover for territories from same study sites.

207

208 3. Results

209

210 Our preliminary vegetation survey showed that the median covers of vegetation layers were
211 significantly lower on dump sites compared to control sites (Figure 2). In addition, dominant
212 woody plant genera were mostly represented by hardy colonizer species reliant on anemochory.
213 This was in contrast with control sites, where typical forest species dominated (Supplementary
214 material, Table S3, Figure S1). Similarly, our analysis of invertebrate community showed
215 significant differences in invertebrate food supply between dump sites and control areas. Orders
216 such as Diptera or Hymenoptera were more typically found on dump sites, while control sites
217 contained more Orthoptera or Aranea (Supplementary material, Table S3, Figure S2). However,
218 most invertebrate orders had higher abundances on control sites rather than dump sites (8 out of
219 12).

220 Our main survey of bird community revealed important differences between dump and
221 control sites. During the survey, a total of 1293 individuals out of 58 bird species have been
222 recorded (Supplementary material, Table S1). The most abundant bird was the chiffchaff
223 (*Phylloscopus collybita*, N = 168), followed by Eurasian blackcap (*Sylvia atricapilla*, N = 92)
224 and the European robin (*Erithacus rubecula*, N = 89). However, our Mann-Whitney U-tests
225 showed that both species richness and bird abundances were significantly higher on control sites

226 compared to dump sites (Figure 3). This means that most bird individuals were found on control
227 sites (N = 799) compared to dump sites (N = 494). The differences between habitat types were
228 also reflected by our CCA analysis of community structure (Table 1, Figure 4). Some species
229 showed clear preference for dump sites (e.g., white wagtail *Motacilla alba*, black redstart
230 *Phoenicurus ochruros*), while others were more common on control areas (e.g., Eurasian nuthatch
231 *Sitta europaea*, Long-toed treecreeper *Certhia familiaris*).

232 However, even species which occurred commonly on both dump and control sites showed
233 very different ecological behaviour between habitat types. Our analysis of habitat preferences
234 showed significant differences in vegetation structure within the 50-meter radii around their
235 territory centres (Table 2). For example, herb layer cover in the territories of five species was
236 significantly lower on dump sites compared to control sites (blue tit *Cyanistes caeruleus*, great
237 tit *Parus major*, willow warbler *Phylloscopus trochilus*, Eurasian blackbird *Turdus merula*,
238 European robin and yellowhammer *Emberiza citrinella*, Figure 5a). In addition, understory cover
239 also differed significantly between habitat types for territories of four species (great tit, chiffchaff,
240 Eurasian blackbird, chaffinch *Fringilla coelebs*), being lower on dump sites (Figure 5b). Canopy
241 cover showed a similar trend in eight species (blue tit, great tit, willow warbler, chiffchaff,
242 Eurasian blackcap, Eurasian blackbird, chaffinch, corn bunting, Figure 5c). These results show
243 that individual birds had to adapt to the hostile environment on the heaps by relaxing their habitat
244 requirements.

245 4. Discussion

246
247 Our findings on differences in vegetation characteristics followed the trends outlined by other
248 ecological studies on dump sites with unmodified surface areas (Bardaghi et al. 2023).
249 Functionally, most dominant woody species recorded on dump sites are representatives of early
250 anemochorous colonizers capable of surviving under tough environmental conditions
251 (Hendrychová et al. 2009). Transition to competitively strong woody genera that are typically
252 present during late successional stages on poor substrates (e.g., *Quercus* sp., *Pinus* sp., Chytrý
253 2017) has not yet proceeded within the 26 to 65-year period of vegetation succession. Exposure
254 to erosion, temperature extremes and unsuitable soil conditions along with toxic elements are
255 likely causes of delayed or arrested succession of vegetation (Novák & Prach 2003). The
256 specificity of our study sites can also stem from accumulation of radionuclides in individual
257 plants, especially in woody genera (Apps et al. 1988). This accumulation may further affect bird

258 abundances since other biota consuming plants with radioactive elements, such as rodents and
259 invertebrates, are hunted by birds (Cleveland et al. 2021). Invertivory is a key strategy for bird
260 species within most climatic zones (Tobias et al. 2022). Thus, the suitability of dump sites for
261 bird inhabitants may be connected with the abundance of invertebrates that serve as food supply.
262 Our results showed that most invertebrate orders sought out control sites rather than dump sites,
263 which might be caused by an impoverished food supply at dump sites. However, our study is not
264 sufficiently exhaustive in its sampling methodology to fully encompass the variation of
265 invertebrate community. A more in-depth examination was conducted by Peterková (2021) who
266 found that invertebrate abundance was contingent on development of woody vegetation. Dump
267 sites with increased afforestation rate were similar in their invertebrate community to the
268 surrounding landscape. However, because vegetation cover is overall less developed on dump
269 sites compared to control sites, it is likely that there is also a decreased availability of
270 invertebrates for birds to exploit.

271 The decreased vegetation cover and insufficient food supply are then likely some of the
272 proximate causes for the relatively low bird species richness and abundances on dump sites. Some
273 comparisons can be made with other studies on birds on early-successional stages of post-mining
274 sites. Šálek (2012) has examined birds of opencast coal mining pits. He found that early
275 successional stages of unreclaimed areas characterized by high proportion of bare ground
276 included the same bird species as dump sites within our study (for example little plover
277 *Charadrius dubius* or white wagtail). In contrast, the bird species he found on later afforested
278 stages were more abundant on control sites within our study (blackcap, chiffchaff). During the
279 survey on another large, reclaimed surface mining site (Korejs et al. 2023) found some species
280 that preferred grassland-based habitats. These species showed preference for dump sites within
281 our study (e.g. yellowhammer or skylark *Alauda arvensis*). It appears that the lack of vegetation
282 cover on the dump sites facilitated suitable habitat for species that either rely on bare ground or
283 require only a semi-intact herb layer. This may enhance the conservation value of dump sites for
284 birds, as groups of species are endangered by habitat loss due to anthropogenic interference due
285 to removal of early-successional habitats (Korejs et al. 2023) or agricultural intensification (Reif
286 & Vermouzek 2019). However, our results also showed that even species that preferred dump
287 sites over the surrounding landscape were forced to modify their habitat requirements (e.g.,
288 yellowhammer used over 30% less herb cover within its dump site territories). The relative
289 avoidance of dump sites by forest and shrub species, such as the blackcap or chiffchaff, is also
290 not common for areas that have undergone several decades of succession. For example,

291 Hendrychová et al. (2009) and Šálek (2012) found that even without assisted revegetation, these
292 species sought out later successional stages on mining sites and spoil heaps. This difference calls
293 back to the observed staggered succession of vegetation (Thrippleton et al. 2018) as well as
294 decreased availability of invertebrates, as many shrub and forest bird species in Central Europe
295 are opportunistic or obligatory invertivores (Tobias et al. 2022). The changes in habitat
296 preferences also suggest increased stress stemming from insufficient proportions of understory
297 and canopy cover that may drive avoidance of dump sites by these otherwise common species
298 (Rosenzweig 1991).

299 We have found that despite being species-poor and low-abundance, bird communities on
300 the dump sites are not without conservation value. The dump sites offer surrogate habitat for
301 primary succession specialists or declining farmland species, but they also function as perching
302 sites for some raptors (Eurasian hobby *Falco subbuteo*), and foraging areas for nomadic
303 granivores (common linnet *Carduelis cannabina*). Aside from birds, they were also found to
304 contain several endangered species of spiders and carabid beetles (Peterková 2021) as well as
305 rare species of poor substrate herbs (dropwort *Filipendula vulgaris*, alpine willowherb *Epilobium*
306 *dodonaei*, observed by authors). The current plant cover on the dump sites has also been found
307 capable of effectively assisting in phytoremediation (Unterbrunner et al. 2007). It is likely that a
308 full-scale restoration effort involving dump site deconstruction, surface levelling and subsequent
309 addition of fertile soils planted with regular rows of trees, would increase overall diversity of
310 birds (Korejs et al. 2023). However, this process would also involve disappearance of endangered
311 habitat specialists that are responsible for the bulk of conservation value on these dump sites, as
312 a similar development was observed on surface mining sites that underwent technical reclamation
313 (Tropek et al. 2012). It should also be noted that the financial costs of disassembling and
314 relocating hundreds of thousands of cubic meters of accumulated mine tailing would be
315 exorbitant (Prach & Hobbs 2008). Therefore, we suggest that the present containment of uranium
316 mine tailing and slag within dump sites is probably most appropriate for biodiversity conservation
317 as well as from an economic point of view.

318 While our results show significant changes in bird community structure and even habitat
319 preferences, the underlying effects responsible for these changes are highly interconnected. Each
320 dump site contains residual uranium together with previously described extreme conditions
321 (Peterková 2021). As such, we cannot single out the effect of radioactivity without taking into
322 consideration other drivers of extreme conditions. Moreover, dump sites around Příbram
323 represent unique habitats, and there is no similar site within the Czech Republic replicating such

324 environmental conditions without comparable history of uranium mining (Mihalík et al. 2011).
325 To test specifically the effects of radioactivity, individual birds would need to be captured and
326 examined for phenotypic effects correlated with genetic damage (Moller et al. 2013). It would be
327 also necessary to examine the radionuclide content within bird tissues (Krivolutski et al. 1999),
328 as other toxic elements associated with underground mining, such as lead or cadmium, may cause
329 physiological damage (Galhardi et al. 2020).

330 **5. Conclusions**

331

332 Rather than attempting to isolate drivers of bird community changes, this study aims to give a
333 general overview of a unique habitat created by uranium mining for birds as a study group. The
334 characteristics of these dump sites prevent full remediation and associated surface modification,
335 which is typical on post-mining sites within Central Europe. However, the examinations of
336 present biotic communities reveal that this may be for the best, as they are unique and worthy of
337 ecological conservation. Our study can serve as a pilot survey giving a first look at bird
338 communities on dump sites after uranium mining, as well as a template for explorations of similar
339 habitats. This study also supplies background information necessary for researching effects of
340 radioactivity by showing how conditions on dump sites alter bird communities. The most suitable
341 expansion of investigated topics could be based on in-depth analysis of phenotypical signs of
342 radiation-associated damage of bird individuals. This could possibly provide a more in-depth
343 look at the drivers of comparatively low abundance and species richness on the dump sites
344 compared to control sites.

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346

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500

501 **7. Tables**502 **Table 1**

503 The effect of location (dump site or control area) on bird community structure (represented by
 504 individual species' abundances), calculated using CCA analysis (see methods).

505

Dependent variable	Predictor	Adjusted explained variation (fraction)	Pseudo-F	p
Community structure	Location (Dump/Control)	0.049	38.18	0.032

506

507 **Table 2**

508 The effect of location (dump site or control area) on habitat preferences of bird species on
 509 examined sites. Dependent variables representing habitat preferences are proportional covers of
 510 different vegetation layers within the 50 m buffer around the estimated centre of individual birds'
 511 territories (see methods). Random effects were represented by 1) site identity, grouping
 512 observations from same dump sites or control areas and by 2) bird species, grouping observations
 513 from the territory of birds from the same species.

514

Dependent variable	Predictor	Estimate	SE	z	p
Herb cover	Location (Dump/Control)	0.706	0.171	38.18	<0.001
Understory cover	Location (Dump/Control)	102.420	47.470	2.158	0.030
Canopy cover	Location (Dump Control)	267.530	84.620	3.162	<0.001

515

516

517

518 8. Figure captions

519 **Figure 1**

520 A map of our study sites and the surrounding landscape with field survey design. Highlighted
521 are dotted lines delimiting the 50 m distance from either the border of the dump sites (black fill)
522 or the border of the control areas (grey fill). The dump sites are also marked with numbers
523 referring to their official ID in the Supplementary material, Table S1, where more information
524 on their parameters is available, taken from DIAMO, state enterprise (2023). Acquired from
525 ArcGIS Pro basemap service (Esri 2023). Mapped data were acquired from
526 <https://geoportal.cuzk.cz/>.

527 **Figure 2**

528 **a)** Photography of typical vegetation cover on the dump sites (left) or the control areas (right). **b)**
529 Comparison of the proportional covers of different vegetation layers on the dump sites or control
530 areas. Mann-Whitney test results are displayed. Thick lines – median, boxes – 25-75% of data,
531 whiskers – non-outlier range.

532 **Figure 3**

533 Boxplots comparing **a)** bird species richness and **b)** bird abundances on dump sites and control
534 areas. Mann-Whitney test results are displayed. Thick lines – median, boxes – 25-75% of data,
535 whiskers – non-outlier range.

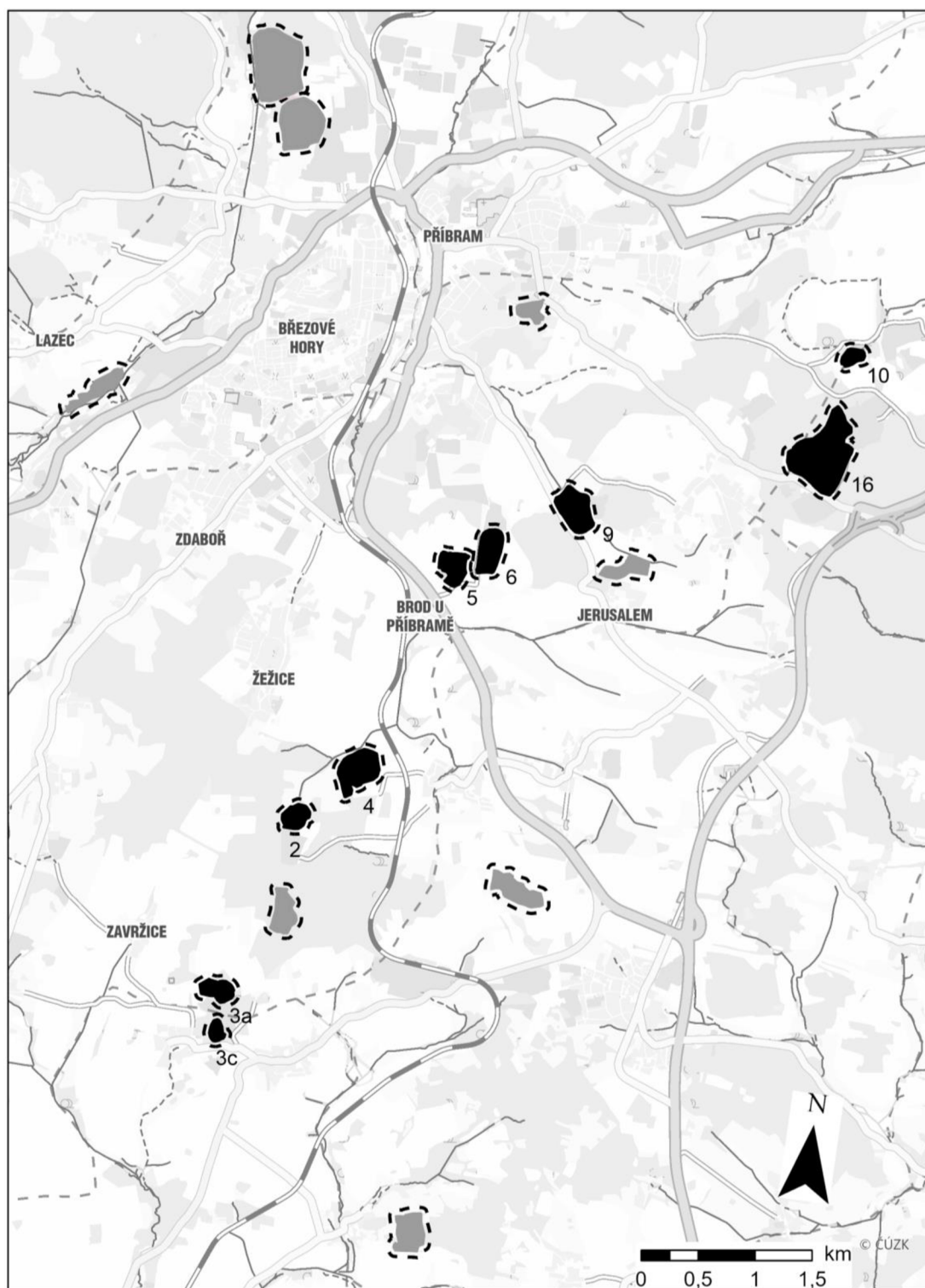
536 **Figure 4**

537 **a)** Ordination diagram comparing bird community composition between dump sites and control
538 areas. Displayed are 19 best fitting bird species. For a list explaining species name abbreviations,
539 see Supplementary material, Table S1. The first and second ordination axes explained together
540 34% of variation.

541 **Figure 5**

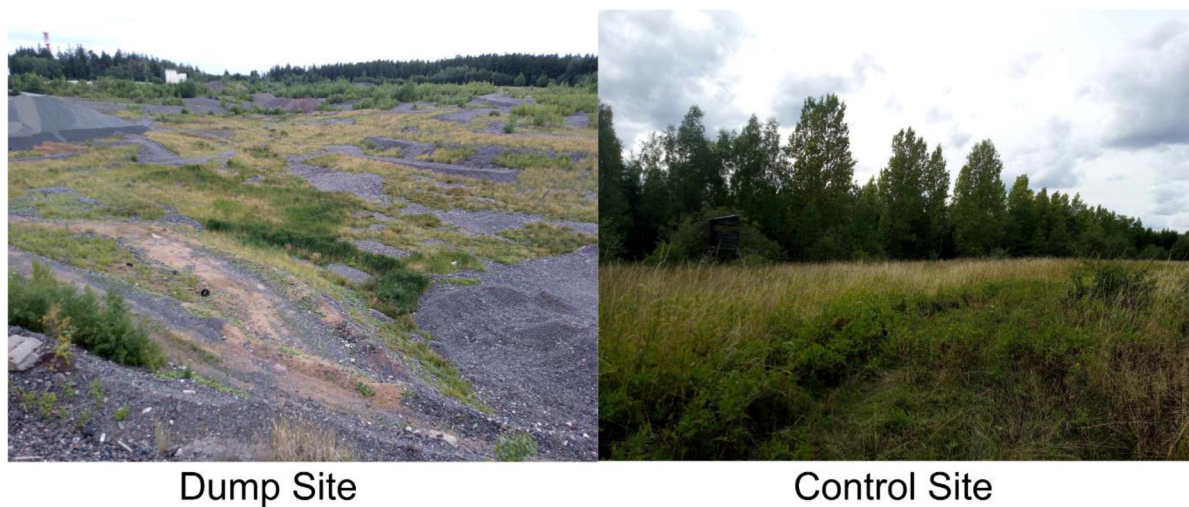
542 Proportional covers **a)** the herb layer **b)** the understory layer and **c)** the canopy layer within the
543 50 m buffer around the estimated centre of the territories of the nine common bird species,
544 displayed for both dump sites and control areas. Added are results of pairwise comparisons using
545 the emmeans function (Lenth et al. 2018). Thick lines – median, boxes – 25-75% of data, whiskers
546 – non-outlier range.

547 9. Figures

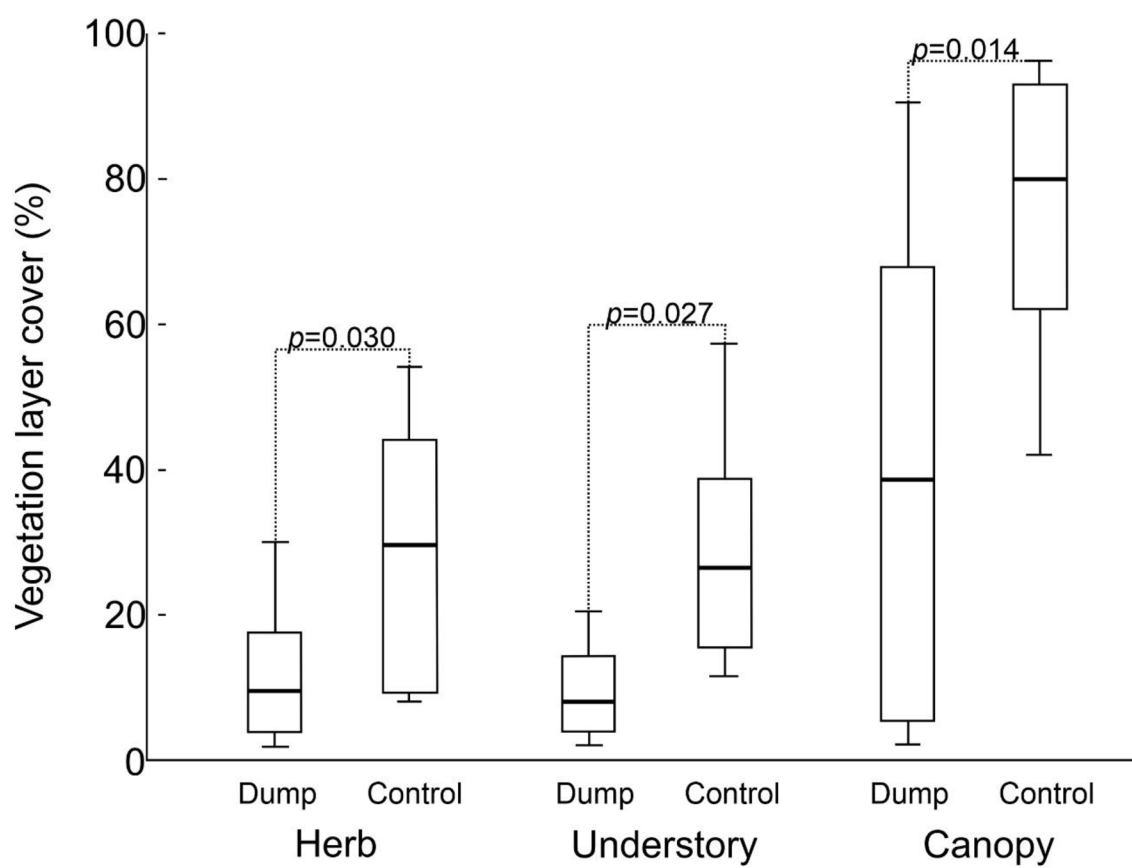
548 **Figure 1**

550 **Figure 2**

a)



b)



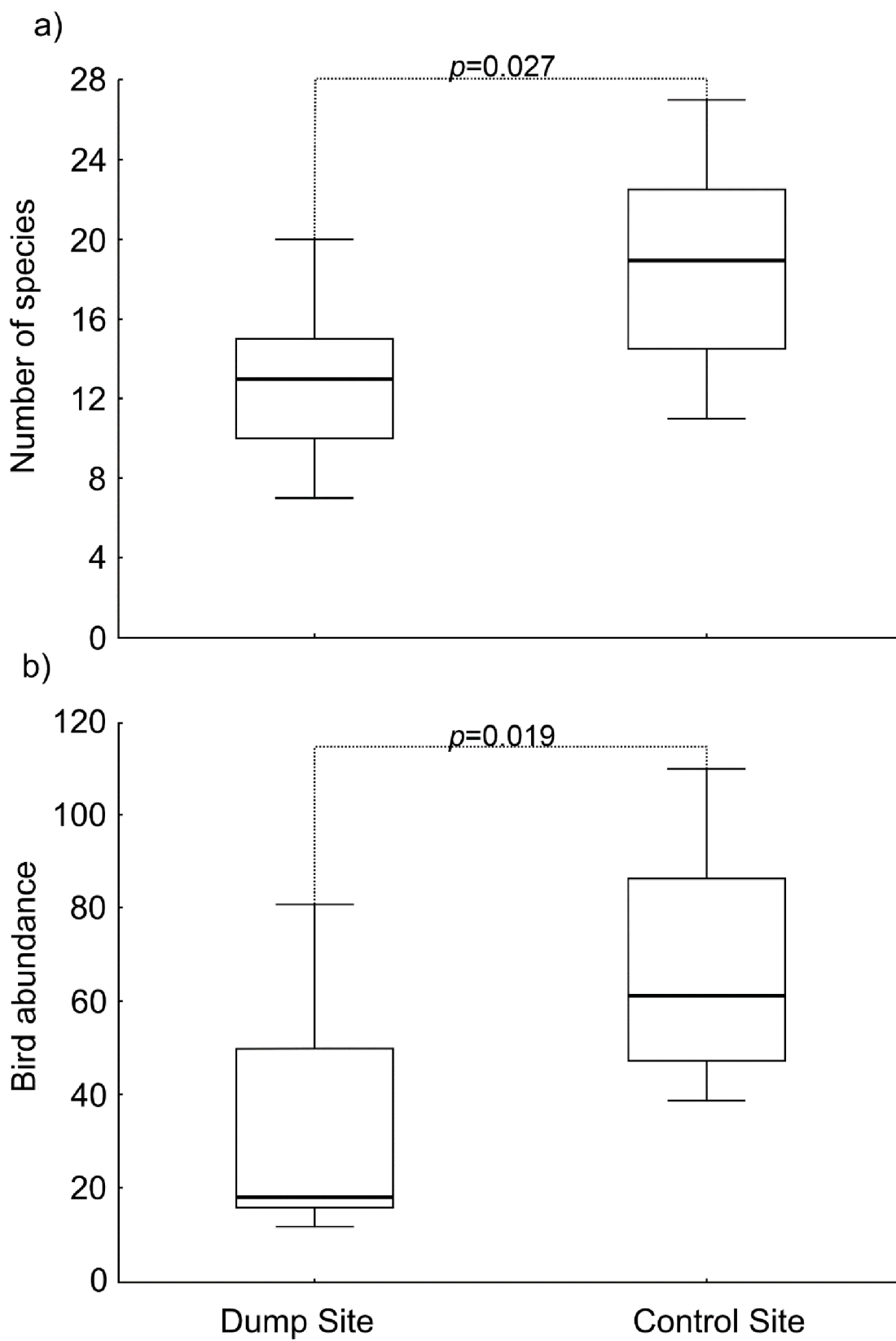
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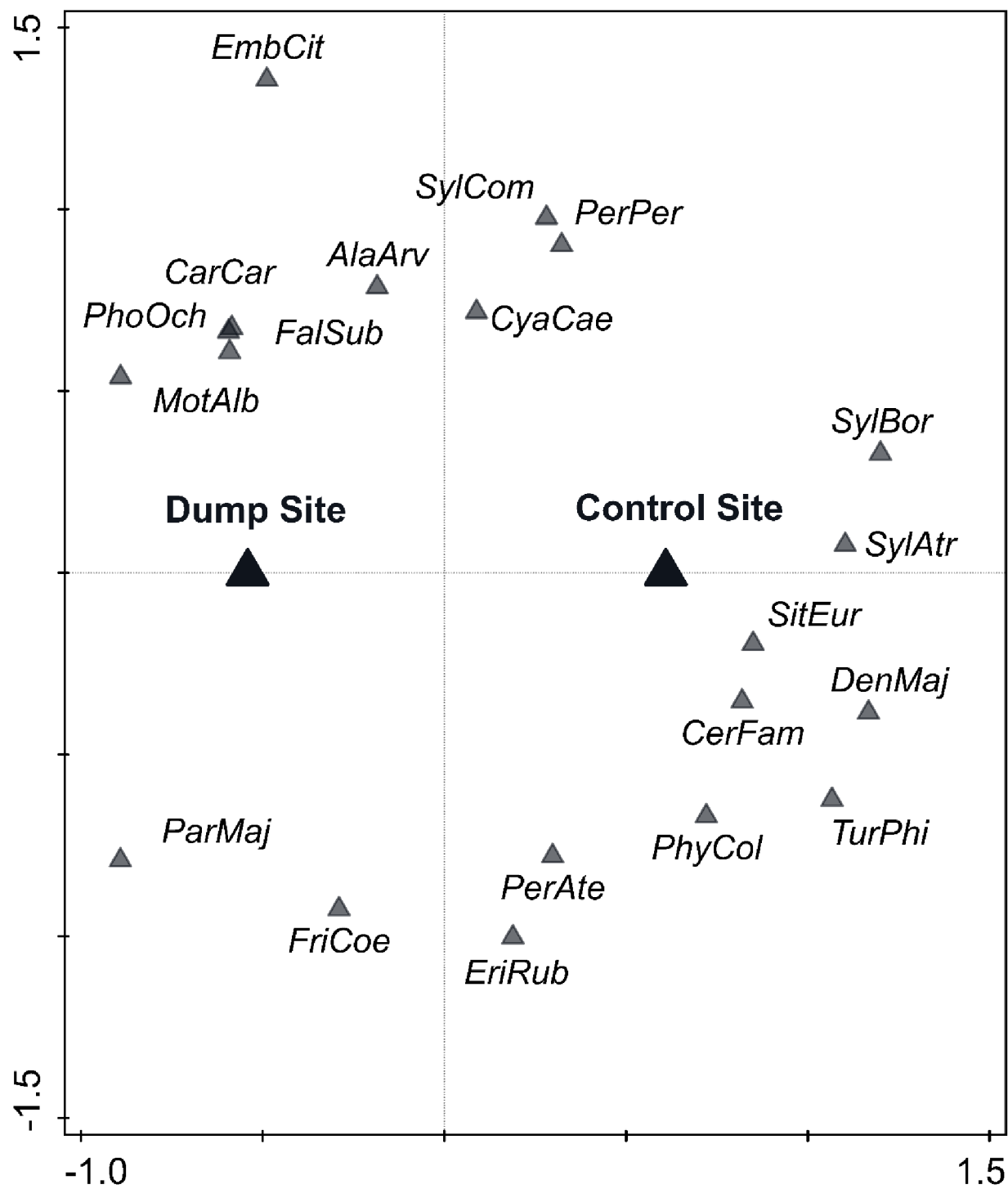
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555 **Figure 3**



557 Figure 4



559 **Figure 5**

